ORIGINAL RESEARCH PAPER Received: 2012.03.09 Accepted: 2012.09.14 Published electronically: 2012.10.01 Acta Soc Bot Pol 81(3):159-165 D0I: 10.5586/asbp.2012.028

Macroalgae fouling community as quality element for the evaluation of the ecological status in Vela Luka Bay, Croatia

Gorana Jelic Mrcelic^{1*}, Merica Sliskovic¹, Boris Antolic²

¹ Maritime Faculty, University of Split, Zrinsko Frankopanska 38, 21 000 Split, Croatia

² Institute of Oceanography and Fisheries, Setaliste Ivana Mestrovica 63, 21 000 Split, Croatia

Abstract

One year qualitative and quantitative study of communities of three major taxonomic groups has been carried out at test panles placed in the upper infarlittoral zone of coastal area of Vela Luka Bay, Croatia. A list of 44 taxa was recorded. *Chaetomorpha* sp., *Ulva* sp., *Fosliella farinosa*, *Sphacelaria cirrosa*, *Polysiphonia scopulorum* were the most frequent dominant taxa. Among 27 algal taxa with noticeable presence only three were classified as ESG (Ecological State Groups) I. Low diversity and species richness together with massive presence of the green algae (as Ulva sp.) and negligible presence of ESG I taxa, may lead to erroneous conclusion that Vela Luka Bay is eutrophicated area. Low values of biomass and *R/P* (Rhodophyceae by Phaeophyceae ratio) Index together with dominance of Phaeophyta also support conclusion that there is no negative impact of nutrient enrichment on macrophyta fouling community in Vela Luka Bay.

Keywords: artificial habitats, bioindicators, coastal waters, fouling organisms, Vela Luka Bay

Intoduction

The major structuring factors influencing the benthic communities are species recruitment onto a surface, competition between resident organisms and disturbance by predation and/ or environmental factors [1], although pollution also influences the development of these communities [2]. Changes in the development of benthic communities caused by organic pollution are often obscured by the interactions between nutrient enrichment and a variety of other ecological factor [3]. In order to distinguish these effects, the use of artificial panels is a good alternative method to field studies [4]. The panels may be easily manipulated and placed under a variety of environmental conditions [1]. Artificial structures are colonized by the most competitive assemblages of floral and fauna species in response to a combination of physical, chemical and biological factors, from the intertidal to the shallow subtidal [5]. Number of factors (age, texture, depth, complexity, inclination and position in the water column) influences colonization of epibiota on artificial panels [6–10]. Vertical surfaces are more densely colonized than horizontal surface [8,11,12].

This is an Open Access digital version of the article distributed under the terms of the Creative Commons Attribution 3.0 License (creativecommons.org/licenses/by/3.0/), which permits redistribution, commercial and non-commercial, provided that the article is properly cited.

In the last years, the increasing need for stable and comparable criteria of environmental quality in the European aquatic ecosystems, reactivated the use and search of pollution biological indicators [13]. Some ecological indices are focused on the presence/absence of a given indicator species, while others take into account species biomass and abundance, the different ecological strategies adopted by organisms [Feldmans *R/P* (Rhodophyceae by Phaeophyceae ratio) index] [14], the diversity (species richness), or the energy variation in the system through changes in the biomass of individuals [13]. Evidence on the suitability of benthic macrophytes as indicators of effects against different pollution gradients is undoubted [15,16]. Marine benthic macrophytes, as photosynthetic sessile organisms being at the base of food web, are vulnerable and adaptive to human and environmental stress of water and sediment. They respond to aquatic environment representing reliable indicators of its changes [17]. A universal pattern is that highly stressed or disturbed marine environments are inhabited by annual species with high growth rates and reproductive potential, while undisturbed marine environments by perennial species with low growth rates and reproductive potential [18-21]. This was the spark to develop the biotic index EEI (Ecological Evaluation Index), based on the functional-morphological model of Littler and Littler [22] and use it to divide marine benthic macrophytes in two different ecological groups, the late-successional [perennials, ESG (Ecological State Groups) I] and the opportunistic (annuals, ESG II) [23]. Thus, the presence and/or the abundance of some benthic macrophytes could be used for the classification of the Ecological (quality) Status, in the terms of the Water Frame Directive [(WFD) 2000/ 60/ EC) [24]. A cost-effective monitoring system to cover the demands of WFD could include summer destructive samplings,

^{*} Corresponding author. Email: gjelic@pfst.hr

functional-form classification, and use of the EEI. Colonization by local infra-upperlittoral benthic communities can be used as biological indicator of environmental changes because they are exhaustively studied [25] as well as they integrate the environmental changes occurring in marine ecosystems, and they are strongly affected by pollution [26]. Further studies are needed to better understand periphyton response to different substrata types and possible seasonal changes of the community structure, especially studies on pollution impact on the macroalgal communities of artificial substrata [27]. The aim of the present study was to give the development of fouling communities using artificial panels and to demonstrate that several ecological indexes based on the composition and abundance of the test panels phytobenthos in the upper infralittoral zone could give good tools for the rapid assessment of the ecological quality of coastal waters.

Material and methods

The fouling community on polyester test panels was studied in the Vela Luka Bay - Bobovišće, the island of Korcula, the Adriatic Sea (42°50' N and 16°43' E). The Vela Luka Bay is closed, shallow and hidden bay, well protected from the north and the south winds. According to Dadić et al. [28], high water oscillations are predominant (in several days periods) during the summer. During the winter, low frequency oscillations are predominant and water exchange is not altered, e.g. water enters in surface layer and exits in bottom layer. Time needed for water exchange ranges from one day to five days depending on sea conditions, wind effects, currents and the stratification in the bay. Water exchange time is longer during the summer when contaminant inflow is the highest. Although the bay is closed and shallow, the oxygen content is high. Water quality analysis show high nitrate, nitrite and phosphorus content, especially in shallow regions of the bay. Phytoplankton structure and biomass shows that Vela Luka Bay is naturally eutrophic area. According to zooplankton biomass this bay can be categorized as the zone D (high zooplankton biomass).

The test-panels were placed in the area of shipyard "Greben" (site Bobovišće) in order to evaluate the environmental impact caused by urbanisation, tourisms and shipbuilding industry.

Four samplings were carried out every three mounths in period between July and July (throughout one year), in order to monitor the characteristics of fouling vegetation on polyester test panels (20×20 cm). One metal structure was placed at the quay, and eight test panels were fixed vertically by ropes, at approximately 0.5 m and 2.5 m depth. Panels were collected after 3, 6, 9 and 12 months of immersion. This overall design gave information on succession patterns and potential recruit available at different period of the year. After collection of the panels were kept in the buckets containing 2% formaldehyde for further analysis.

Examination of the panels took place in The Laboratory of Benthos – The Institute of Oceanography and Fishery, Split. The panels were carefully removed from the buckets in the laboratory, placed in an aquarium filled with seawater, and photographed with a digital camera. According to Boduresque [29] and Braun-Blanquet [30] total cover percentage of fouling community was determined for each panel. The structure of the fouling community was estimated using the Constant Area Method. Four subsamples (2 × 2 cm squares) were randomly scratched from each side of the panel, i.e. eight samples per

Tab. 1 The list of algal taxa found on test panels at both depths.

CHLOROPHYTA

Bryopsidophyceae

Blastophysa sp. Blastophysa rhizopus Reinke Bryopsis sp.

Dasycladophyceae

Acetabularia acetabulum (Linnaeus) Silva

Ulvophyceae

Cladophora sp. Chaetomorpha sp. Chaetomorpha aerea (Dillwyn) Kuetzing Ulothrix sp. Enteromorpha sp. Ulva sp. Ulvella sp. Phaeophila dendroides (Crouan) Batters

РНАЕОРНҮТА

Phaeophyceae

Dictyota dichotoma (Hudson) Lamouroux Dictyota linnearis (C. Agardh) Gerville Padina pavonica (Linnaeus) Thivy Feldmannia irregularis (Kuetzing) Hamel Colpomenia sinuosa (Mertens) Derbes i Solier Ectocarpus paradoxus Montagne Ectocarpus confervoides Kjellmann Entonema sp. Myrionema sp. Myriotrichia sp. Giraudia sphacelarioides Derbes & Solier Sphacelaria cirrosa (Roth) C. Agardh Sphacelaria plumula Zanardini Sphacelaria tribuloides Meneghini Stypocaulon scoparium (Linnaeus) Kuetzing

RHODOPHYTA

Bangiophyceae

Goniotrichum alsidii (Zanardini) Howe

Compsopogonophyceae

Erythrotrichia carnea (Dillywyn) J. Agardh

Florideophyceae

Fosliella farinosa (Lamouroux) Howe Acrochaetium sp. Acrochaetium davesii (Dillwyn) Naegeli Aglaothamnion furcellarie (J. Agardh) G. Feldmann Antithamnion cruciatum (C. Agardh) Naegeli var. profundum G. Feldmann Ceramium codii (Richards) G. Mazoyer Heterosiphonia wurdemannii (Bailey) Falkemberg Dasya ocellata (Grateloup) Harvey Dasya arbuscula (Dillwyn) C. Agardh Lophosiphonia cristata Falkemberg Polysiphonia scopulorum Harvey Laurencia sp. Gelidium sp.

Tab. 1 (continued)

Falkenbergia rufolanosa (Harvey) Schmitz

panel. Samples were examined by light microscopy (Carl Zeiss – Jena, ocular 10×, lens 8 and 40×). Each sample was carefully sorted and identification at species and functional group-level was attempted. Fouling species were identified using authoritative keys and texts [31–36].

Phytobenthos fouling community structure was analyzed in terms of total coverage, species number, total species coverage, species frequency, Sörensen similarity coefficient [37] qualitative dominance (DN%) of main algal systematic groups, R/Pratio and classification in ESG groups, in order to get an indication of the state of health of the study area. The total percentage covers of species (Ri) and mean cover of species (RM) were analyzed according to Boudouresque [29] and Braun-Blanquet et al. [30]. The frequency and classification in ESG groups were estimated only for algal taxa with noticeable presence (>1% cover). Shifts in marine ecosystem structure and function are evaluated by classifying marine benthic macrophytes in two ESGs (I, II), representing alternative ecological states, e.g. pristine and degraded [17]. The total wet weight of fouling organisms (TW, g) was weighted on Tecnica - type EXACTA 1200 EB precision electronic scale (within 0.0001g), after the panel were completely scratched.

Results

Fouling biomass ranged from 55 g × m⁻² (after 3 month of immersion at 2.5 m depth) to 965.25 g × m⁻² (after 12 month of immersion at 2.5 m depth). Number of algal taxa ranged from 4 to 18 and the number of animal taxa ranged from 2 to 11. Number of algal taxa was considerably higher than the number of animal taxa (except after 6 month of immersion at 0.5 m depth). Total number of algal taxa was 44 (37 taxa at 0.5 m depth and 41 taxa at 2.5 m depth) – Tab. 1. That gave 34 algal species common for both depths, i.e. Sörensen similarity coefficient was 87.18%.

The *R/P* ranged from 0.33 to 1.20 at 0.5 m depth, and from 0.33 to 1.67 at 2.5 m depth. At 0.5 m depth Phaeophyta (except at the inner side of the panel after 6 month and after 12 month) were dominant over Rhodophyta and Chlorophyta with minimum 12.50% and maximum 50%. Qualitative dominance of Chlorophyta ranged between 20% and 42%, while qualitative dominance of Rhodophyta ranged between 12.50% and 40%. At 2.5 m depth there were no dominant algal groups. Qualitative dominance of Chlorophyta ranged between 20% and 40%, while qualitative dominance of Phaeophyta ranged with minimum 23% and maximum 60%, and qualitative dominance of Rhodophyta ranged between 20% and 44%. Fig. 1 and Fig. 2 show the qualitative dominance (*DN*%) of taxa of the main algal groups found on test panels at two different depths.

Chaetomorpha sp. and Polysiphonia scopulorum (100%); Ulva sp. and Fosliella farinosa (87.50%) and Sphacelaria cirrosa (81.25%) were the most frequent taxa. Among 27 algal taxa with noticeable presence only three were classified as ESG I (Acetabularia acetabulum, Padina pavonica and Fosliella farinosa) – Tab. 2.

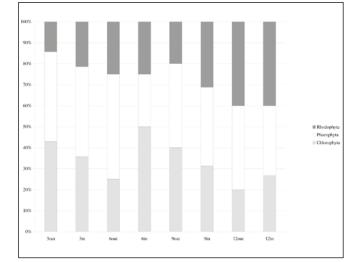


Fig. 1 The qualitative dominance (*DN*%) of taxa of the main algal groups found on test panels (in – inner side; out – outer side) at 0.5 m.

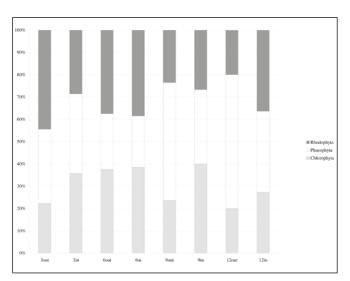


Fig. 2 The qualitative dominance (*DN*%) of taxa of the main algal groups found on test panels (in – inner side; out – outer side) at 2.5 m.

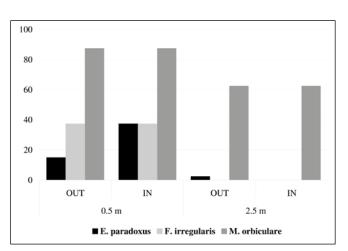


Fig. 3 The mean cover (%) of the dominat algal taxa found on test panels (in – inner side; out – outer side) at 0.5 m and 2.5 m after 3 months of immersion.

Tab. 2 Ecological State Groups classification and the frequency of algal taxa covering at least 1% of the sampling area.

Таха	ESG	Frquency (f/16)
Acetabularia acetabulum (Linnaeus) Silva	Ι	1
Cladophora sp.	II	2
<i>Chaetomorpha</i> sp.	II	16
Pheophila dendroides (Crouan) Batters	II	7
Rhizoclonium sp.	II	12
Ulotrix sp.	II	4
Ulva sp.	II	14
Dictyota dichotoma (Hudson) Lamouroux	II	7
Dictyota linnearis (C. Agardh) Gerville	II	1
Ectocarpus paradoxus Monatgne	II	7
Entonema sp.	II	5
Feldmannia irregularis (Kuetzing) Hamel	II	9
Giraudya sphacelarioides Derbes i Solier		12
Myrionema orbiculare J. Ag.	II	12
Padina pavonica (Linnaeus) Thivy	Ι	5
Sphacelaria cirrosa (Roth) C. Agardh	II	13
Sphacelaria plumula Zanardini	II	2
Sphacelaria tribuloides Meneghini	II	3
Aglaothamnion furcellarie (J. Agardh) G. Feldmann	II	3
Antithamnion cruciatum (C. Agardh) Naegeli var.	II	7
<i>profundum</i> G. Feldmann		
Dasya arbuscula (Dillwyn) C. Agardh	II	2
Fosliella farinosa (Lamouroux) Howe	Ι	14
Heterosiphonia wurdemannii (Bailey) Falkemberg	II	1
Lophosiphonia cristata Falkemberg	II	9
Polysiphonia scopulorum Harv.	II	16

ESG - Ecological State Groups.

Acetabularia acetabulum had frequency 6.25% and RM of 3% (r), Padina pavonica had frequency 31.25% and RM raged between 0% and 15% (form + to 2), while Fosliella farinosa had frequency 87.50% and mean cover (RM) raged between 0% and 15% (form + to 2).

After 3 months of immersion, total percentage cover was 90% (5 in Braun-Blanquet [30] scale) at 0.5 m at both sides of the panels; while total percentage cover was 90% (5) at the outer side and 70% (4) at the inner side of the panel at 2.5 m. *Myrionema orbiculare, Ectocarpus paradoxus* and *Feldmannia irregularis* were dominat algal taxa after 3 months of immersion (Fig. 3).

After 6 months of immersion, total percentage cover was 100% (5) at both depths and both sides of the panels. *Sphacelaria cirrosa* and *Polysiphonia scopulorum* were dominat algal taxa after 6 months of immersion (Fig. 4).

After 9 months of immersion, total percentage cover was 90% (5) at both depths and both sides of the panels. *Cladophora* sp., *Giraudya sphacelarioides*, *Myrionema orbiculare* and *Polysiphonia scopulorum* were dominat algal taxa after 9 months of immersion (Fig. 5).

After 12 months of immersion, total percentage cover was 100% (5) at both depths and both sides of the panels. *Cladophora* sp., *Ulva* sp. and *Polysiphonia scopulorum* were dominat algal taxa after 12 months of immersion (Fig. 6).

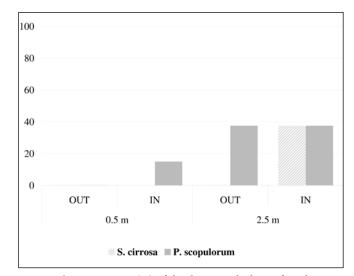


Fig. 4 The mean cover (%) of the dominat algal taxa found on test panels (in – inner side; out – outer side) at 0.5 m and 2.5 m after 6 months of immersion.

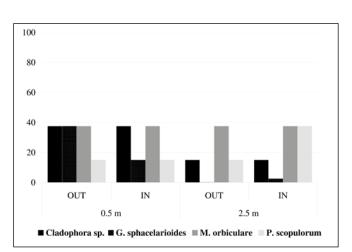


Fig. 5 The mean cover (%) of the dominat algal taxa found on test panels (in – inner side; out – outer side) at 0.5 m and 2.5 m after 9 months of immersion.

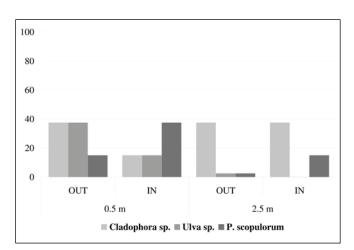


Fig. 6 The mean cover (%) of the dominat algal taxa found on test panels (in – inner side; out – outer side) at 0.5 m and 2.5 m after 12 months of immersion.

Discussion

Organic pollution may increase and speed up the development of fouling communities in originally oligotrophic area [3], or at higher levels, may cause a decrease of biomass and diversity and favor opportunistic species [19,38]. Benthic macrophyta community directly responds to the changes in abiotic variables, reducing its richness, diversity levels and/ or enhancing its biomass per surface unit under eutrophication [39]. Therefore it can be used as sensitive bioindicator of ecosystem changes at smaller spatial scales [27]. They provide readable responses with analysis based on a functional group level providing powerful support to traditional species-level analysis. Benthic macrophyte communities have been successfully used as indicators of eutrophication in coastal waters of The Adriatic Sea [40–42].

Fouling biomass was relatively low compare to a similar 12 month study of fouling communities on concrete, plastic and glass panels in Kastela Bay (The Adriatic Sea; from 221 g × m⁻² to 61.578 g × m⁻²) [43].

Number of found algal taxa were considerably lower compared to 62 benthic algae taxa found on concrete, plastic and glass panels at the entrance of Kastela Bay [43]. Low diversity and species richness may be the result of organic enrichment [27], althought some authors suggest that low diversity is caused by competition between tolerant and non-tolerant species [3].

In polluted areas, the number of algae species decreases, especially Rhodophytes and Phaeophytes, with an increase in the abundance of Chlorophytes [19,44]. Phaeophyceae are extremly sensitive to environmental disturbance [27]. Feldman's *R*/*P* Index [14], based on marine vegetation, is highly used in the Mediterranean Sea. It was established as a biogeographical index and it is based on the fact that the number of species of Rodophyceae decreases from the Tropics to the Poles. Its application as indicator holds on the higher or lower sensitivity to disturbances of Phaeophyceae and Rhodophyceae [13]. Low R/P values and dominance of Phaeophyta over Rhodophyta and Chlorophyta at 0.5 m depth indicated that there was no negative impact of nutrient enrichment on macrophyta. In some cases functional groups was preferable than taxonomic grouping of organisms to reduce spatial and temporal community variability and to discover patterns without loosing important information [23]. Among 27 algal taxa with noticeable presence only three were classified as ESG I (Acetabularia acetabulum, Padina pavonica and Fosliella farinosa). The reason for low presence of late successional ESG I taxa could be the short period of immersion (less than one year).

Although found dominant taxa are very adaptive and represent typical fouler in the Adriatic [45], numerous authors pointed out that macrolagae of genus *Ulva* [46–48] and Cladophora [49] appears in conditions with high nutrient loading [26,46]. The massive presence of the green algae *Ulva* sp. and *Cladophora* sp. along the European coastline is considered to be a reliable indicator for nutrient enriched seawater [50]. Chlorophytes are favoured by an increase of nutrients, and genera such as *Cladophora* and *Ulva* are usually abundant in eutrophicated areas due to their great reproductive capacity and their rapid growth rate [44].

Low diversity and species richness together with massive presence of the green algae (as *Ulva* sp.) and negligible presence of ESG I taxa, may lead to erroneous conclusion that Vela Luka Bay is eutrophicated area. Nevertheless, similar situation

Conclusions

(*i*) Low biomass, diversity, species richness, *R/P* Index, dominance of Phaeophyta, together with massive presence of green algae (as *Ulva* sp.) and negligible presence of ESG I taxa can result from nutrient enrichment in Vela Luka Bay, but also from short immersion period.

(*ii*) A possibility of wider application of the presence and the abundance of benthic macrophytes for the classification of the Ecological Status is shown.

(*iii*) The analysis of macrophyta fouling communities (functional-form classification, EEI) on artificial test panels obtains simple, cost-effective, non-destructive monitoring system that covers the demands of Water Frame Directive (2000/ 60/ EC).

References

- Richmond MD, Seed R. A review of marine macrofouling communities with special reference to animal fouling. Biofouling. 1991;3(2):151–168. http://dx.doi. org/10.1080/08927019109378169
- Meyer-Reil LA, Köster M. Eutrophication of marine waters: effects on benthic microbial communities. Mar Pollut Bull. 2000;41(1-6):255-263. http://dx.doi.org/10.1016/ S0025-326X(00)00114-4
- Mayer-Pinto M, Junqueira AOR. Effects of organic pollution on the initial development of fouling communities in a tropical bay, Brazil. Mar Pollut Bull. 2003;46(11):1495– 1503. http://dx.doi.org/10.1016/S0025-326X(03)00249-2
- Rastetter EB, Cooke WJ. Responses of marine fouling communities to sewage abatement in Kaneohe Bay, Oahu, Hawaii. Mar Biol. 1979;53(3):271–280. http://dx.doi. org/10.1007/BF00952436
- Juanes JA, Guinda X, Puente A, Revilla JA. Macroalgae, a suitable indicator of the ecological status of coastal rocky communities in the NE Atlantic. Ecol Indic. 2008;8(4):351– 359. http://dx.doi.org/10.1016/j.ecolind.2007.04.005
- Connell SD, Glasby TM. Do urban structures influence local abundance and diversity of subtidal epibiota? A case study from Sydney Harbour, Australia. Mar Environ Res. 1999;47(4):373–387. http://dx.doi.org/10.1016/ S0141-1136(98)00126-3
- Connell SD. Urban structures as marine habitats: an experimental comparison of the composition and abundance of subtidal epibiota among pilings, pontoons and rocky reefs. Mar Environ Res. 2001;52(2):115–125. http://dx.doi.org/10.1016/S0141-1136(00)00266-X
- Glasby TM, Connell SD. Orientation and position of substrata have large effects on epibiotic assemblages. Mar Ecol Prog Ser. 2001;214:127–135. http://dx.doi.org/10.3354/ meps214127
- 9. Svane I, Petersen JK. On the problems of epibioses, fouling and artificial reefs, a review. Mar Ecol. 2001;22(3):169–188. http://dx.doi.org/10.1046/j.1439-0485.2001.01729.x
- 10. Perkol-Finkel S, Shashar N, Benayahu Y. Can artificial reefs

mimic natural reef communities? The roles of structural features and age. Mar Environ Res. 2006;61(2):121–135. http://dx.doi.org/10.1016/j.marenvres.2005.08.001

- Langhamer O, Wilhelmsson D, Engström J. Artificial reef effect and fouling impacts on offshore wave power foundations and buoys – a pilot study. Estuar Coast Shelf Sci. 2009;82(3):426–432. http://dx.doi.org/10.1016/j. ecss.2009.02.009
- Knott NA, Underwood AJ, Chapman MG, Glasby TM. Epibiota on vertical and on horizontal surfaces on natural reefs and on artificial structures. J Mar Biol Assoc UK. 2004;84(6):1117–1130. http://dx.doi.org/10.1017/ S0025315404010550h
- Salas F, Marcos C, Neto JM, Patrício J, Pérez-Ruzafa A, Marques JC. User-friendly guide for using benthic ecological indicators in coastal and marine quality assessment. Ocean Coast Manage. 2006;49(5–6):308–331. http://dx.doi. org/10.1016/j.ocecoaman.2006.03.001
- 14. Pérez Ruzafa I. Efecto de la contaminación sobre la vegetación submarina y su valor indicador. In: Pérez Ruzafa Á, Marcos C, Salas F, Zamora S, editors. Perspectivas y herramientas en el estudio de la contaminación marina. Murcia: Servicio de Publicaciones de la Universidad de Murcia; 2003. p. 133–147.
- 15. Chryssovergis F, Panayotidis P. Evolution des peuplements macrophytobenthiques le long d'un gradient d'eutrophisation. Oceanol Acta. 1995;18:649–658.
- 16. Schiel DR, Wood SA, Dunmore RA, Taylor DI. Sediment on rocky intertidal reefs: Effects on early post-settlement stages of habitat-forming seaweeds. J Exp Mar Biol Ecol. 2006;331(2):158–172. http://dx.doi.org/10.1016/j. jembe.2005.10.015
- 17. Orfanidis S, Papathanasiou V, Sabetta L, Pinna M, Gigi V, Gounaris S, et al. Benthic macrophyte communities as bioindicators of transitional and coastal waters: relevant approaches and tools. Transit Water Bull. 2007;1(3):45–49. http://dx.doi.org/10.1285/i1825229Xv1n3p45
- Regier HA, Cowell EB. Applications of ecosystem theory, succession, diversity, stability, stress and conservation. Biol Conserv. 1972;4(2):83–88. http://dx.doi. org/10.1016/0006-3207(72)90002-X
- Murray SN, Littler MM. Patterns of algal succession in a perturbated marine intertidal community. J Phycol. 1978;14(4):506-512. http://dx.doi.org/10.1111/j.1529-8817.1978.tb02477.x
- 20. Sousa WP. The responses of a community to disturbance: The importance of successional age and species' life histories. Oecologia. 1980;45(1):72–81. http://dx.doi. org/10.1007/BF00346709
- 21. Schramm W. Factors influencing seaweed responses to eutrophication: some results from EU-project EUMAC. J Appl Phycol. 1999;11(1):69–78. http://dx.doi. org/10.1023/A:1008076026792
- 22. Littler MM, Littler DS. The evolution of thallus form and survival strategies in benthic marine macroalgae: field and laboratory tests of a functional form model. Am Nat. 1980;116(1):25. http://dx.doi.org/10.1086/283610
- 23. Panayotidis P, Montesanto B, Orfanidis S. Phytobenthos as a quality element for the ecological status evaluation: a case study of the implementation of the Water Frame Directive (2000/ 60/ EC) in the mediterranean ecoregion. In: Proceedings of the Second Mediterranean Symposium on Marine Vegetation. Athens: UNEP/MAP/RAC/SPA;

2003. p. 211–216.

- 24. UNEP/MAP. Fact sheets on marine pollution indicators [UNEP(DEC)/MED WG.264/Inf.14]. Athens: UNEP; 2005.
- 25. Angel D. An application of artificial reefs to reduce organic enrichment caused by net-cage fish farming: preliminary results. ICES J Mar Sci. 2002;59 suppl:324–329. http:// dx.doi.org/10.1006/jmsc.2002.1208
- 26. Torras X, Pinedo S, Garcia M, Mangialajo L, Ballesteros E. Assessment of coastal environmental qualita based on littoral community cartography: methodological approach. In: Proceedings of the Second Mediterranean Symposium on Marine Vegetation. Athens: UNEP/MAP/RAC/SPA; 2003. p. 145–153.
- 27. Mannino AM, Sara G. Effects of fish-farm biodeposition on periphyton assemblages on artificial substrates in the southern Tyrrhenian Sea (Gulf of Castellammare, Sicily). Aquat Ecol. 2007;42(4):575–581. http://dx.doi.org/10.1007/ s10452-007-9131-1
- 28. Dadić V, Gačić M, Galasso I, Glavurtić M, Krstulović N, Marasović I, et al. Istrazivanje fizikalnih karakteristika i prijemnog kapaciteta zaljeva Vela Luka s posebnim osvrtom na izgradnju marine. Split: Institute of Oceanography and Fisheries; 1991.
- 29. Boudouresque CF. Methodea detude qualitative et quantitative du benthos (en paticulier du phytobenthos). Tethys. 1971;3(1):79–104.
- Braun-Blanquet J, Fuller GD, Conard HS. Plant sociology; the study of plant communities. London: Hafner Pub. Co.; 1965.
- Coppejans E. Iconographie d'Algues Mediterraneene. Vaduz: Bibliotheca Phycologica; 1983.
- 32. Costello MJ, Emblow CS, White R. European register of marine species. Aix: Patrimoines Naturels; 2001.
- Feldmann J. Les algues marines de la côte des Albères. I-III. Cyanophycées, Chlorophycées, Phaéophycées. Rev Algol. 1937;9:149–335.
- Occhipinti-Ambrogi A. Contributo alla conoscenza dei Briozoi nella Laguna Veneta settentrionale. Boll Mus civ St Nat Venezia. 1981;31:95–109.
- 35. Parenzan P. Carta d'identita delle conchigllie del Mediterraneo. Bivalvii parte seconda. Tarnto: Bios-Taras; 1976.
- 36. Riedl R. Fauna e flora del Mediterraneo. Padova: Franco Muzzio Editore; 1991.
- 37. Sørensen TJ. A method of establishing groups of equal amplitude in plant sociology based on similarity of species content and its application to analyses of the vegetation on danish commons. Kgl Dan Vid Selsk Biol Skr. 1948;4:1–34.
- 38. Hargrave BT, Thiel H. Assessment of pollutioninduced changes in benthic community structure. Mar Pollut Bull. 1983;14(2):41-46. http://dx.doi. org/10.1016/0025-326X(83)90189-3
- 39. Mirto S, La Rosa T, Gambi C, Danovaro R, Mazzola A. Nematode community response to fish-farm impact in the western Mediterranean. Environ Pollut. 2002;116(2):203– 214. http://dx.doi.org/10.1016/S0269-7491(01)00140-3
- 40. Iveša L, Lyons DM, Devescovi M. Assessment of the ecological status of north-eastern Adriatic coastal waters (Istria, Croatia) using macroalgal assemblages for the European Union Water Framework Directive. Aquat Conserv. 2009;19(1):14–23. http://dx.doi.org/10.1002/aqc.964
- 41. Orlando-Bonaca M, Lipej L. Benthic macroalgae is bioindicators of the ecological status in the Gulf of Trieste. Varstvo Narave. 2009;22:63–72.

- 42. Sliskovic M, Jelic-Mrcelic G, Antolic B, Anicic I. The fouling of fish farm cage nets as bioindicator of aquaculture pollution in the Adriatic Sea (Croatia). Environ Monit Assess. 2010;173(1–4):519–532. http://dx.doi.org/10.1007/ s10661-010-1402-y
- 43. Span A, Antolic B, Hell Z. Istrazivanje bioloskog aspekta obrastanja na uronjenim staklenim i plasticnim plocama i betonskim stupicima. Split: Institute of Oceanography and Fisheries; 1990.
- 44. Borowitzka MA. Intertidal algal species diversity and the effect of pollution. Mar Freshwater Res. 1972;23(2):73–84. http://dx.doi.org/10.1071/MF9720073
- Igic LJ. Prikaz obrastaja u Jadranskom i drugim svjetskim morima - biolosko znacenje. Pomorski Zbornik. 1995;33:329–356.
- 46. Golubic S. Effect of organic pollution on benthic communities. Mar Pollut Bull. 1970;1(4):56–57. http://dx.doi. org/10.1016/0025-326X(70)90121-9
- Ballesteros E, Pérez M, Zabala M. Aproximación al conocimiento de las comunidades algales de la zona infralitoral superior en la costa catalana. Collect Bot. 1984;15:69–100.

- Rodríguez-Prieto C, Polo L. Effects of sewage pollution in the structure and dynamics of the community of *Cystoseira mediterranea* (Fucals, Phaephyceae). Sci Mar. 1996;60:253–263.
- Belsher T. Analyse des répercussions des pollutions urbaines sur le macrophytobenthos de la Méditerranée (Marseille, Port-Ventres, Port-Cros). Aix: Paul Cézanne University; 1977.
- 50. Council Directive for a legislative frame and actions for the water policy, 2000/ 60/ EC. Official Journal of the European Communities. 2000.12.22.
- 51. Ibanez F, Dauvin JC. Long-term changes (1977 to 1987) in a muddy fine sand *Abra alba-Melinna palmata* community from the Western English Channel: multivariate time-series analysis. Mar Ecol Prog Ser. 1988;49:65–81. http://dx.doi. org/10.3354/meps049065
- Craeymeersch JA. Applicability of the abundance/biomass comparison method to detect pollution effects on intertidal macrobenthic communities. Hydrobiol Bull. 1991;24(2):133–140. http://dx.doi.org/10.1007/BF02260430