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Demographic population structure and fungal associations of plants colonizing High Arctic glacier forelands, Petuniabukta, Svalbard

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Abstract

The development of vegetation in Arctic glacier forelands has been described as unidirectional, non-replacement succession characterized by the gradual establishment of species typical for mature tundra with no species turnover. Our study focused on two early colonizers of High Arctic glacier forelands: Saxifraga oppositifolia (Saxifragaceae) and Braya purpurascens (Brassicaceae). While the first species is a common generalist also found in mature old growth tundra communities, the second specializes on disturbed substrate. The demographic population structures of the two study species were investigated along four glacier forelands in Petuniabukta, north Billefjorden, in central Spitsbergen, Svalbard. Young plants of both species occurred exclusively on young substrate, implying that soil conditions are favourable for establishment only before soil crusts develop. We show that while S. oppositifolia persists from pioneer successional stages and is characterized by increased size and flowering, B. purpurascens specializes on disturbed young substrate and does not follow the typical unidirectional, non-replacement succession pattern. Plants at two of the forelands were examined for the presence of rootassociated fungi. Fungal genus Olpidium (Fungus incertae sedis) was found along a whole successional gradient in one of the forelands.

Large areas of Arctic and alpine regions have been deglaciated and thus have become open to plant colonization since the end of the Little Ice Age (LIA) about 120 years ago. The boundaries of the last maximum development of glaciers are usually easily distinguished in the field (Rachlewicz et al. 2007), making glacier forelands an outdoor laboratory for the study of plant strategies in cold regions. Colonization of glacier forelands is hindered by a cold and short growing season, small and variable seed production, rough substratum, scarcity of nutrients and organic matter and disturbance by runoff, solifluction, cryoturbation and permafrost (Svoboda & Henry 1987; Chambers 1995; Hodkinson et al. 2003; Moreau et al. 2005; Moreau et al. 2008; Yoshitake et al. 2010).

Plants that are especially successful in primary succession of deglaciated areas are recruited from a pool of stress-tolerant species of cold regions (Caccianiga et al. 2006). These colonizers are dispersed to the deglaciated areas, where seeds germinate at low temperatures and become established (Marcante et al. 2009; Schweinbacher et al. 2012). In contrast to succession on new substrates in temperate regions (e.g., Řehounková & Prach 2010), the early stages of succession in deglaciated areas of the Arctic lack ruderal species characterized by a



short lifespan and large investment into fecundity (Moreau et al. 2008; Moreau et al. 2009; Prach & Rachlewicz 2012). That is, in the temperate alpine regions, pioneer plants rapidly colonize recently deglaciated areas within a few years and are replaced with late-successional plant assemblages in the following few decades of succession (Whittaker 1993; Raffl et al. 2006). In the Arctic, in contrast, pioneer plants take decades to colonize and the colonizers often persist in old tundra communities (Svoboda & Henry 1987; Hodkinson et al. 2003; Moreau et al. 2008; Mori et al. 2008; Prach & Rachlewicz 2012). Owing to a reduced species pool and environmental conditions, early successional specialist species are sparse in the High Arctic (Svoboda & Henry 1987) and/or they persist in old growth tundra on disturbed microsites, so their successional status is blurred.

Positive biotic interactions-for example, mycorrhizal symbioses (Nakatsubo et al. 2010; Fujiyoshi et al. 2011), amelioration of substrate by cyanobacterial crusts (Yoshitake et al. 2010) and accumulated organic matter (Hodkinson et al. 2003)—play an important role in early stages of the succession in this low energy environment. It has been suggested that facilitation and competition are rare due to low plant cover (Nakatsubo et al. 2010). During succession in Arctic deglaciated areas substrates change over time. Initially they can be relatively nutrient rich, then reach a quasi-steady-state of development during the fourth or fifth decade due to low precipitation, which limits the mineral weathering (Kabala & Zapart 2012). Only after 60 years do soils develop (Hodkinson et al. 2003). In conjunction with this substrate succession, several studies have reported changing diversity and frequency of root-associated fungi (e.g., Jumpponen 2003; Cazarés et al. 2005; Fujiyoshi et al. 2011). While arbuscular mycorrhiza is generally considered rare at higher latitudes (Newsham et al. 2009), being only rarely reported from Svalbard (Väre et al. 1992; Öpik et al. 2013), it has been suggested that diverse functional groups of fungi (e.g., dark septate endophytes) could influence nutrient uptake and growth in Arctic plants (Newsham et al. 2009). Ectomycorrhiza and association with dark septate endophytes have repeatedly been observed in the Arctic regions (Kohn & Stasovski 1990; Väre et al. 1992; Fujimura & Egger 2012).

Most pioneer species in the High Arctic are regarded as generalists capable of colonizing young substrate, in addition to other habitats, resulting in the unidirectional, non-replacement succession pattern (Svoboda & Henry 1987; Prach & Rachlewicz 2012). Here, we argue that in addition to these generalists, some pioneer species are specialists restricted to young substrate. The generalist species can be expected to display an increase of size and proportion of flowering individuals (i.e., an ageing pattern) along the successional gradient of time since deglaciation. By contrast, an early-succession specialist should depend more on local conditions and population structure should be more or less independent of the time elapsed since deglaciation. To identify these patterns, we investigated demographic population structures of two pioneer species-a widely spread generalist, Saxifraga oppositifolia, and an assumed specialist, Braya purpurascens-along the successional gradients of four glacier forelands in the Petuniabukta area, northern Billefjorden, Spitsbergen (Fig. 1). Further, to contribute to our understanding of the role of fungal symbioses in succession in the High Arctic, we examined roots of the two species for the presence of root-associated fungi on two of the forelands.

Materials and methods

Study species

Saxifraga oppositifolia (Saxifragaceae) and Braya purpurascens (Brassicaceae) have been reported as pioneer species in glacier foreland succession in Svalbard (Hodkinson et al. 2003; Moreau et al. 2008; Nakatsubo et al. 2010; Prach & Rachlewicz 2012). Saxifraga oppositifolia holds its position up to late stages of succession and is common in old growth tundra. In contrast, *B. purpurascens* is frequent in initial stages of succession, but its abundance decreases in later stages, becoming limited to disturbed places in old growth tundra (Moreau et al. 2008; Prach et al. 2012).

Saxifraga oppositifolia is a long-lived perennial clonal species, which regenerates both from seeds and vegetative diaspores. In the study area, young plants are prostrate, non-clonal and have a main root. Older plants produce adventitious roots on branches and start to disintegrate from old parts (Klimešová et al. 2012). Branches can be deposited by runoff water and re-root (Hagen 2002). *Saxifraga oppositifolia* produces a large number of tiny, well-dispersed seeds forming a long-term persistent seed bank (Cooper et al. 2004; Schweinbacher et al. 2010; Müller et al. 2012). The species is facultatively mycorrhizal, forming arbuscular mycorrhiza or can be non-mycorrhizal (Harley & Harley 1987).

Saxifraga oppositifolia displays great morphological, genetic and ploidy variability through its circumpolar Arctic and alpine distribution, including Svalbard (Müller et al. 2012). There are two main morphological forms: prostrate and cushion (Kume et al. 1999; Kume 2003). Most plants in this study were growing on young substrate (less



Fig. 1 (a) Map of studied area of Petuniabukta. The studied glacier forelands are marked with dashed lines. The inset shows the location of northern Billefjorden on the island of Spitsbergen. (b) The glacier forelands involved in the study with reconstructed positions of glacier foreheads since the Little Ice Age (LIA) maximum (based on Rachlewicz et al. 2007). The red circles depict positions of the study sites.

than 15–20 years) and were too young and small (Fig. 2) often just a single unbranched shoot—to be classified into a growth form. Nevertheless, the vast majority of older plants (ca. 95%) clearly belonged to the prostrate form, with only five clear cushion-type plants found at sites 2–5 of the Sven foreland.

Braya purpurascens (syn. *Braya glabella* subsp. *purpurascens*) is restricted to the Arctic. It is a non-clonal plant with a perennial main root and one to several leaf rosettes and flowering stems (Klimešová et al. 2012; Fig. 3). Bledsoe et al. (1990) found no fungal colonization on roots for this species.

Data collection

Data on both study species were collected on forelands of four glaciers—Ferdinand, Hørbye, Ragnar and Sven, in Petuniabukta, northern Billefjorden, central Spitsbergen, Svalbard—in late July through early August 2011. A transect was established on each glacier foreland beginning at the closest occurrence of either of the studied species to the glacier front. The transects were ca. 1500 m long in the cases of the Ferdinand, Hørbye, Ragnar forelands, while the transect on the Sven foreland ended on the oldest LIA moraine. Using a global positioning system device, series of plots were placed along each time-since-deglaciation gradient at distances of 100, 300, 700 and 1500 m from the starting point (Fig. 1b). Plants were sampled from central parts of those plots until about 20 plants were sampled. Therefore, the sampled area was affected by the number of plant individuals on a site and plot size varied from about 100 m² to several thousand m². For each foreland, the age since deglaciation was derived from dates of glacier retreat presented by Rachlewicz et al. (2007; Fig. 2).

Up to 20 plants of each species were sampled at each site to determine the demographic structure of the populations. *Saxifraga oppositifolia* was the most frequent species and 20 individuals were sampled at most sites. In contrast, *Braya purpurascens* was in low abundance and rather scattered at most sites. Therefore, the *B. purpurascens* data set does not cover the whole glacier forelands in most cases and less than 20 individuals were sampled at many sites.

The following traits of *S. oppositifolia* individuals were measured: size of plant (product of length of the longest axis of a tuft and that orthogonal to it) and reproductive status of the plant (flowering/fruiting vs. sterile). For *B. purpurascens*, the reproductive status of the plant (flowering/fruiting vs. sterile) and number of sterile and fertile rosettes per individual were counted. The tuft size, number of rosettes or proportion of flowering individuals



Fig. 2 Variation in shoot size and architecture of *Saxifraga oppositifolia* plants growing on the gradient on the Hørbye glacier foreland: (a) seedling with less than 1 cm² cover in the initial succession stage (9 years since deglaciation); (b) a small sterile plant in the initial succession stage (9 years since deglaciation); (c) a more advanced sterile plant with longer shoots (11 years since deglaciation); (d) a loose tuft forming plant with few flowers (13 years since deglaciation); (e) a dense tuft forming flowering plant (18 years since deglaciation).

represent the demography of the population integrating age of plant individuals and variation in growth rates at different sites (Gatsuk et al. 1980).

Roots of five randomly chosen individuals of both species were collected at Ragnar and Hørbye forelands at the youngest inhabited site, at 300 m and 1500 m distant sampling sites. While *S. oppositifolia* individuals occurred at all the sampling sites, *B. purpurascens* was missing from the initial point and the 300 m distant site at Ragnar. Roots were washed, silica gel-dried for transport, and subsequently re-hydrated and stained by Chlorazol Black E following the procedure in Šmilauerová et al. (2012). Ten randomly chosen roots per species and sampling site were inspected by light microscopy.

Statistical analyses

Patterns of demographic population structure (i.e., size and reproductive status of plant individuals) of both study species along the gradients of the age since deglaciation were assessed by a series of generalized linear and non-linear models. Due to different shapes of the gradients on the four forelands, each foreland was analysed separately, though the data and graphical representation of the models were subsequently embedded in single figures. *Saxifraga oppositifolia* size was analysed

by non-linear hyperbolic models with the following equation:

size
$$=$$
 $\frac{a}{b + age} + c,$ (1)

where *a*, *b* and *c* are parameters defining the shape of the relationship and age is the site age since deglaciation as derived from Rachlewicz et al. (2007). The size was log_{10} transformed prior to the analysis. This model assumes a greater size difference between plots in the young part of the age gradient than in the old part and an asymptote to which size will converge with the age. The asymptote is represented by the parameter *c* of the model. In addition, the projected intercept of the hyperbola with the x-axis (i.e., estimated age of the colonization start; backtransformed size $10^0 = 1 \text{ mm}^2$) is fitted by the model and corresponds to $age_0 = -a/c - b$. Reproductive status of both S. oppositifolia and B. purpurascens was analysed by binomial generalized linear models. The demographic population structure of *B. purpurascens*, based on numbers of plants differing in reproductive status (flowering vs. sterile) and number of rosettes (single vs. multiple), was visualized by bar-plots and analysed by goodness-of-fit test as contingency tables where possible. All statistical analyses were conducted in R version 2.13.2 (R Development Core Team 2011).



Fig. 3 Variation in shoot size and architecture of *Braya purpurascens* plants growing on the Hørbye glacier foreland: (a) a sterile plant with two rosettes (9 years since deglaciation); (b) flowering single-rosette (FS) plant (11 years since deglaciation); (c) flowering plant with multiple rosettes and inflorescences (11 years since deglaciation).

Results

While both species occurred on the glacier forelands there was variability in their abundance, shoot architecture and size (Figs. 2, 3). Observed at all 20 sites, *Saxifraga oppositifolia* was confirmed as an omnipresent pioneer on the glacier forelands in Petuniabukta. We detected a clear relationship between the average size of *S. oppositifolia* plants and the time since site deglaciation (Figs. 4, 5, Table 1). Characterized by an initial steep increase of size but reaching a plateau on sites ca. 20 years older that the youngest occurrence (Fig. 5), this pattern was similar among the forelands but there were notable differences in the projected intercepts of model hyperbolas with the age axis, that is, estimated start of the colonization.

The estimated start of colonization varied among the forelands (Table 1). At young sites (9-26 years since deglaciation), S. oppositifolia plants were just simple shoots a few centimetres long. These small seedlings (Fig. 2a; size less than two on the log-scale corresponding to cover ca. 1 cm²) were frequent at most sites younger than 20 years (Fig. 4). Only rarely did these plants also occur at older sites (Ragnar foreland: sites 2 and 3 with two and one individuals, respectively; Sven foreland: site 4 with one individual). Small sterile plants (Fig. 2b; size between two and three on the log-scale, corresponding to cover of ca. $1-10 \text{ cm}^2$) were common. They were most frequent at young sites and were absent at the oldest sites of Ferdinand and Hørbye forelands. Plants of intermediate size (Fig. 2c; size between three and four on the log-scale, corresponding to cover of ca. 10–100 cm²) were common along the whole gradients with the exception of Hørbye, Ragnar and Sven youngest sites. Large plants (Fig. 2d, e; size larger than four on the log-scale corresponding to cover exceeding ca. 100 cm²) were common or prevailing at older sites but were still present in low numbers at most of the youngest sites. The proportion of flowering individuals in the populations increased with age since deglaciation (Table 2). Plants growing on the youngest sites were always sterile, with sparse flowering recorded at slightly older sites (ca. 20 years since deglaciation). Approximately 50% of the plants were flowering at sites older than 50 years.

Braya purpurascens occurred in scattered populations, and was missing from six of the 20 sampling sites and only one individual was found at three sites. As a result the *B. purpurascens* data set is less comprehensive than that for *S. oppositifolia*. Nonetheless, some inferences can be made. Both species were able to colonize young sites. *Braya purpurascens* did not show clear patterns of demographic structure across the foreland gradients (Fig. 6). The youngest cohort of sterile, single rosette plants (Fig. 3a) and the most advanced flowering, multiple rosette (Fig. 3c) cohort co-occurred at most sites (although the flowering multi-rosette plants were missing from the youngest sites of Ferdinand and Hørbye;



Fig. 4 Histograms of Saxifraga oppositifolia size distribution at individual sites on the glacier forelands under study. Each glacier foreland is represented by a row of plots. Reconstructed ages—expressed as years since deglaciation (ysd)—are given for each site.

Fig. 6). There was a positive relationship between the occurrence of flowering and the age since deglaciation on the Ferdinand and Hørbye gradients; see Fig. 6 for the increasing proportion of the flowering multi-rosette plants (see also Table 2).

No arbuscular mycorrhizal structures were detected in the roots of both species. Both were free from infection at the Hørbye foreland. However, roots of both species contained resting and ripe sporangia typical of *Olpidium* (Fig. 7) on the Ragnar foreland. These structures were found at the three sampling sites (including the youngest succession stage) in *S. oppositifolia* roots and also in *B. purpurascens* roots at the oldest site (*B. purpurascens* was absent from the other two sampling sites).

Discussion

The detected patterns in the demography of the widely distributed generalist *Saxifraga oppositifolia* support the expected pattern of increasing size and proportion of flowering individuals along the successional gradients. By contrast, no such trends were found in *Braya purpurascens*. Large flowering individuals of this species could be found at old and young sites (except for the youngest sites). Its scattered distribution on the forelands suggests this short-lived species is influenced by other environmental factors besides age since deglaciation. Together with its absence from mature tundra (our field observations and Moreau et al. 2009), these findings suggest *B. purpurascens* as a distinct specialized pioneer that deviates from the unidirectional, non-replacement succession pattern in the Arctic (Svoboda & Henry 1987; Prach & Rachlewicz 2012).

To the best of our knowledge, there is only one study evaluating size structure of plant population along a successional gradient in an Arctic glacier foreland. Mori et al. (2008) analysed demographic population structure of *Salix arctica* on a successional gradient spanning 35 000 years on Ellesmere Island, in the Canadian Arctic Archipelago. Despite the long time



Fig. 5 Patterns of *Saxifraga oppositifolia* size on the gradients of the age since deglaciation. Filled points represent mean values within study sites; error bars represent one standard error. See Table 1 for summaries of non-linear models describing the relationship between *S. oppositifolia* size and the years elapsed since deglaciation.

span of succession, the vegetation was very sparse, and small individuals of *Salix arctica* prevailed along the whole gradient despite increasing density and mean size of the individuals. Another study with which we can compare our results comes from an alpine area in southern Norway (Whittaker 1993). The author analysed size structure of six pioneer species and four late-successional shrubs. All but one showed ageing populations, where size correlated with age since deglaciation. Smallest plants prevailed in young successional stages with abundance decreasing, or plants were completely absent with increasing competition in later successional stages. The only exception was *Oxyria digyna*, a snow-bed specialist showing a variable size structure due to its special habitat demands (Whittaker 1993).

Our sites in Svalbard have higher precipitation than Ellesmere Island, but are drier and have shorter growing seasons than southern Norway (Serreze & Barry 2005). Therefore, it would be anticipated that successional processes in Svalbard would be shorter than Ellesmere and longer than southern Norway. Indeed, succession to a closed tundra community is estimated to take about 2000 years in Svalbard (Hodkinson et al. 2003). Although the glacier forelands in our study were characterized by sparse vegetation, the species studied were establishing successfully on new substrate whereas on older substrate, we did not find new seedlings (*S. oppositifolia*) or they were restricted to disturbed sites (*B. purpurascens*). Hence, the new substrate seems more beneficial for the establishment of these species than the old substrate, although the seed bank and seed rain increase with age (Cooper et al. 2004).

Young substrate in Arctic glacier forelands is characterized by high pH and lack of soil crust (Anderson et al. 2000); the studied species are therefore either supported in establishment by basic soil reaction or hindered by soil crust. Cyanobacterial crust is formed within 16 years of succession, when small individual *S. oppositifolia* and

Table 1 Summary of non-linear hyperbolic models describing the patterns in size of *Saxifraga oppositifolia* individuals on the gradients of the age since deglaciation on the individual glacier forelands. The hyperbola parameter *c* corresponds to the asymptote to which size converges with the age; $age_0 (= -a/c - b)$ is the projected intercept between the hyperbolas and the x axis, which corresponds to an estimate of the start of the colonization.

	Нуре	erbola par						
Glacier	а	b	С	age ₀	r ²	F	df	р
Ferdinand	-27.68	4.57	5.16	0.8	0.23	9.54	3.97	< 10 ⁻⁴
Hørbye	-7.77	-5.61	4.55	7.3	0.52	35.61	3.97	< 10 ⁻⁶
Ragnar	-4.8	-13.86	4.15	15.0	0.50	32.74	3.97	< 10 ⁻⁴
Sven	-7.06	-23.00^{a}	4.11	24.7	0.56	41.28	3.97	< 10 ⁻⁶

 $^{\rm a}\text{For}$ this parameter, a minimum limit of -23.00 was set manually to prevent an unrealistic behaviour of the model.

Table 2 Summary of generalized linear models describing proportions of flowering individuals in *Saxifraga oppositifolia* and *Braya purpurascens* populations on the gradients of the age since deglaciation on individual glacier forelands. *P*-values are based on comparisons of the deviance values with a χ^2 distribution with one degree of freedom.

	Sc	axifraga opp	ositifolia	Braya purpurascens				
Glacier	Explained deviance (%)	Deviance	Regression slope	р	Explained deviance (%)	Deviance	Regression slope	р
Ferdinand	9.1	12.20	0.023	< 0.001	9.8	5.88	0.058	0.015 ^a
Hørbye	10.7	10.74	0.060	< 0.002	6.5	6.37	0.010	0.012
Ragnar	10.7	10.42	0.029	< 0.002	Non-significant ($p = 0.13$)			
Sven	13.9	19.27	0.038	$< 10^{-4}$	Non-significant ($p = 0.25$)			

^aThe generalized linear model for *B. purpurascens* on the Ferdinand foreland is affected by an occurrence of just a single flowering plant at the oldest site of the foreland (Fig. 6). If this plant is omitted from the analysis, the result is still significant (explained deviance = 7.2%, regression slope = 0.057, *p* = 0.040).

B. purpurascens have become established (as observed by: Hodkinson et al. 2003, Moreau et al. 2005; Moreau et al. 2008; Prach & Rachlewicz 2012). We could there-

fore speculate that the two studied species need a fresh substrate without soil crust for seed establishment whereas later colonizing species may be unaffected or



Fig. 6 Bar-plots displaying population structure of *Braya purpurascens* at individual sites on the glacier forelands under study. Each glacier foreland is represented by a row of plots. *Braya purpurascens* plants were classified to following classes: sterile single rosette (SS), sterile multiple rosettes (SM), flowering single rosette (FS) and flowering multiple rosettes (FM). Reconstructed ages—expressed as years since deglaciation (ysd)—are given for each site. Contingency table-based tests of differences in population structure between sites within forelands: Ferdinand: $\chi^2 = 17.4$, df = 4, p < 0.002 (sites 1, 2 and 4 were only considered, FS category was omitted due to absence at site), Hørbye: $\chi^2 = 17.7$, df = 9, p < 0.05 (sites 2–4 were considered), Ragnar not tested due to representative population occurring only at site 5, Sven: $\chi^2 = 3.7$, df = 6, p = 0.71 (sites 2–4 were considered).



Fig. 7 Fungal structures of *Olpidium* in roots of *Saxifraga oppositifolia* sampled at (a) Ragnar 1 (closest to the glacier foreland) and (b) Ragnar 3 sampling sites, and (c) *Braya purpurascens* at the Ragnar 5 site. The scale bars are 100 μm.

positively affected by soil crust. Diverse effects of soil crusts on seedling establishment are documented in the literature from warm desert conditions (Godinez-Alvarez et al. 2012) but only a positive effect is reported or expected from cold desert sites (Bliss & Gold 1999; Cooper 2011). Soil crust changes environmental conditions by increasing concentration of carbon, nitrogen, moisture and temperature, decreasing pH, and preventing soil churning in comparison with crust-free soil (Gold 1998; Hodkinson et al. 2003). Other studies, however, report cooling of the soil surface due to the soil crust and no effect of increasing nutrient availability on vascular plants (Gold et al. 2001). It is apparent that further research on soil crust ecology and vegetation development, along the lines of Yoshitake et al. (2010) and Moreau et al. (2005), in the pioneer stages of succession in Arctic glacier forelands is required.

Symbiotic associations with mycorrhizal fungi and dark septate endophytes are important for plants growing in nutrient-poor ecosystems (e.g., Newsham et al. 2009). However, they are perhaps not essential given the absence of these fungal mutualists in the roots of these pioneer species growing on recently deglaciated soils. Instead, structures of zoosporic fungal genus *Olpidium* comprising biotrophic pathogens were found. The genus *Olpidium*, formerly placed among Chytridiomycota, now has an uncertain phylogenetic position and is likely related to former Zygomycota (Sekimoto et al. 2011). Our finding add to evidence of the frequent occurrence of zoosporic fungi (including former Chytridiomycota) in high elevation or recently deglaciated Arctic soils (e.g., Jumpponen 2003; Schmidt et al. 2012).

Our study shows that *B. purpurascens* specializes on new or disturbed substrate lacking soil crust. While *S. oppositifolia* also establishes itself on new substrate, because of its longevity it persists in old growth tundra. These results imply that different plant strategies in primary

succession can be recognized in the High Arctic, although they are probably only represented by a few species.

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References

- Anderson S.P., Drever J.I., Frost C.D. & Holden P. 2000. Chemical weathering in the foreland of a retreating glacier. *Geochimica et Cosmochimica Acta 64*, 1173–1189.
- Bledsoe C., Klein P. & Bliss L.C. 1990. A survey of mycorrhizal plants on Truelove Lowland, Devon Island, N.W.T., Canada. *Canadian Journal of Botany 68*, 1848–1856.
- Bliss L.C. & Gold W.G. 1999. Vascular plant reproduction, establishment, and growth and the effects of cryptogamic crusts within a polar desert ecosystem, Devon Island, NWT, Canada. *Canadian Journal of Botany* 77, 623–636.
- Caccianiga M., Luzzaro A., Pierce S., Ceriani R.M. & Cerabolini B. 2006. The functional basis of a primary succession resolved by CSR classification. *Oikos 112*, 10–20.
- Cazarés E., Trappe J.M. & Jumpponen A. 2005. Mycorrhiza– plant colonization patterns on a subalpine glacier forefront as a model system of primary succession. *Mycorrhiza 15*, 405–416.
- Chambers J.C. 1995. Disturbance, life history strategies, and seed fates in alpine herbfield communities. *American Journal of Botany 82*, 421–433.
- Cooper E.J. 2011. Polar desert vegetation and plant recruitment in Murchisonfjord, Nordaustlandet, Svalbard. *Geografiska Annaler Series A*, 243–252.

- Cooper E.J., Alsos I.G., Hagen D., Smith F.M., Coulson S.J. & Hodkinson I.D. 2004. Plant recruitment in the High Arctic: seed bank and seedling emergence on Svalbard. *Journal of Vegetation Science 15*, 115–124.
- Fujimura K.E. & Egger K.N. 2012. Host plant and environment influence community assembly of High Arctic root-associated fungal communities. *Fungal Ecology 5*, 409–418.
- Fujiyoshi M., Yoshitake S., Watanabe K., Murota K., Tsuchiya Y., Uchida M. & Nakatsubo T. 2011. Successional changes in ectomycorrhizal fungi associated with the polar willow *Salix polaris* in a deglaciated area in the High Arctic, Svalbard. *Polar Biology* 34, 667–673.
- Gatsuk L.E., Smirnova O.V., Vorontzova L.I., Zaugolnova L.B. & Zhukova L.A. 1980. Age states of plants of various growth forms—a review. *Journal of Ecology 68*, 675–696.
- Godinez-Alvarez H., Morin C. & Rivera-Aguilar V. 2012. Germination, survival and growth of three vascular plants on biological soil crusts from a Mexican tropical desert. *Plant Biology* 14, 157–162.
- Gold W.G. 1998. The influence of cryptogamic crusts on the thermal environment and temperature relations of plants in a High Arctic polar desert, Devon Island, NWT, Canada. *Arctic and Alpine Research 30*, 108–120.
- Gold W.G., Glew K.A. & Dickson L.G. 2001. Functional influences of cryptobiotic surface crusts in an alpine tundra basin of the Olympic Mountains, Washington, USA. *Northwest Science 75*, 315–326.
- Hagen D. 2002. Propagation of native Arctic and alpine species with a restoration potential. *Polar Research 21*, 37–47.
- Harley J.L. & Harley E.L. 1987. A check-list of mycorrhiza in the British flora. *New Phytologist 105*, 1–102.
- Hodkinson I.D., Coulson S.J. & Webb N.R. 2003. Community assembly along proglacial chronosequences in the High Arctic: vegetation and soil development in north-west Svalbard. *Journal of Ecology 91*, 651–663.
- Jumpponen A. 2003. Soil fungal community assembly in a primary successional glacier forefront ecosystem as inferred from analyses of rDNA sequences. *New Phytologist 158*, 569–578.
- Kabala C. & Zapart J. 2012. Initial soil development and carbon accumulation on moraines of the rapidly retreating Werenskiold Glacier, SW Spitsbergen, Svalbard Archipelago. *Geoderma 175*, 9–20.
- Klimešová J., Doležal J., Prach K. & Košnar J. 2012. Clonal growth forms in Arctic plants and their habitat preferences: a case study from Petuniabukta, Spitsbergen. *Polish Polar Research* 33, 421–442.
- Kohn L.M. & Stasovski E. 1990. The mycorrhizal status of plants at Alexandra Fiord, Ellesmere Island, Canada, a High Arctic site. *Mycologia 82*, 23–35.
- Kume A., Bekku Y.S., Hanba Y.T. & Kanda H. 2003. Carbon isotope discrimination in diverging growth forms of *Saxifraga oppositifolia* in different successional stages in a High Arctic glacier foreland. *Arctic, Antarctic, and Alpine Research 35*, 377–383.
- Kume A., Nakatsubo T., Bekku Y. & Masuzawa T. 1999. Ecological significance of different growth forms of purple

saxifrage, *Saxifraga oppositifolia* L., in the High Arctic, Ny-Ålesund, Svalbard. *Arctic, Antarctic, and Alpine Research* 31, 27–33.

- Marcante S., Winkler E. & Erschbamer B. 2009. Population dynamics along a primary succession gradient: do alpine species fit into demographic succession theory? *Annals of Botany 103*, 1129–1143.
- Moreau M., Laffly D. & Brossard T. 2009. Recent spatial development of Svalbard strandflat vegetation over a period of 31 years. *Polar Research 28*, 364–375.
- Moreau M., Laffly D., Joly D. & Brossard T. 2005. Analysis of plant colonization on an Arctic moraine since the end of the Little Ice Age using remotely sensed data and a Bayesian approach. *Remote Sensing of Environment 99*, 244–253.
- Moreau M., Mercier D., Laffly D. & Roussel E. 2008. Impacts of recent paraglacial dynamics on plant colonization: a case study on Midtre Lovénbreen foreland, Spitsbergen (79°N). *Geomorphology 95*, 48–60.
- Mori A.S., Osono T., Uchida M. & Kanda H. 2008. Changes in the structure and heterogeneity of vegetation and microsite environments with the chronosequence of primary succession on a glacier foreland in Ellesmere Island, High Arctic Canada. *Ecological Research* 23, 363–370.
- Müller E., Bronken-Eidesen P., Ehrich D. & Alsos I.G. 2012. Frequency of local, regional and long-distance dispersal of diploid and tetraploid *Saxifraga oppositifolia* (Saxifragaceae) to Arctic glacier forelands. *American Journal of Botany 99*, 459–471.
- Nakatsubo T., Fujiyoshi M., Yoshitake S., Koizumi H. & Uchida M. 2010. Colonization of the polar willow *Salix polaris* on the early stage of succession after glacier retreat in the High Arctic, Ny-Ålesund, Svalbard. *Polar Research 29*, 385–390.
- Newsham K.K., Upson R. & Read D.J. 2009. Mycorrhizas and dark septate root endophytes in polar regions. *Fungal Ecology 2*, 10–20.
- Öpik M., Zobel M., Cantero J.J., Davison J., Facelli J.M., Hiiesalu I., Jairus T., Kalwij J.M., Koorem K., Leal M.E., Liira J., Metsis M., Neshataeva V., Paal J., Phosri C., Polme S., Reier U., Saks U., Schimann H., Thiery O., Vasar M. & Moora M. 2013. Global sampling of plant roots expands the described molecular diversity of arbuscular mycorrhizal fungi. *Mycorrhiza 23*, 411–430.
- Prach K., Klimešová J., Košnar J., Redchenko O. & Hais M. 2012. Variability of contemporary vegetation around Petuniabukta, central Svalbard. *Polish Polar Research 33*, 383–394.
- Prach K. & Rachlewicz G. 2012. Succession of vascular plants in front of retreating glaciers in central Spitsbergen. *Polish Polar Research* 33, 319–328.
- Rachlewicz G., Szcucinski W. & Ewertowski M. 2007. Post-"Little Ice Age" retreat rates of glaciers around Billefjorden in central Spitsbergen, Svalbard. *Polar Biology 28*, 159–186.
- Raffl C., Mallaun M., Mayer R. & Erschbamer B. 2006. Vegetation succession pattern and diversity changes in a glacier valley, central Alps, Austria. *Arctic, Antarctic, and Alpine Research 38*, 421–428.

- R Development Core Team. 2011. *R: a language and environment for statistical computing.* Vienna: R Foundation for Statistical Computing.
- Řehounková K. & Prach K. 2010. Life-history traits and habitat preferences of colonizing plant species in long-term spontaneous succession in abandoned gravel sand pits. *Basic* and Applied Ecology 11, 45–53.
- Schmidt S. K., Naff C. S. & Lynch R. C. 2012. Fungal communities at the edge: ecological lessons from high alpine fungi. *Fungal Ecology* 5, 443–452.
- Schweinbacher E., Navarro-Cano J.A., Neuner G. & Erschbamer B. 2012. Correspondence of seed traits with niche position in glacier foreland succession. *Plant Ecology* 213, 371–382.
- Sekimoto S., Rochon D., Long J.E., Dee J.M. & Berbee M.L. 2011. A multigene phylogeny of *Olpidium* and its implications for early fungal evolution. *BMC Evolutionary Biology* 11, 331.
- Serreze M.C. & Barry R.G. 2005. *The Arctic climate system*. Cambridge: Cambridge University Press.

- Šmilauerová M., Lokvencová M. & Šmilauer P. 2012. Fertilization and forb: graminoid ratio affect arbuscular mycorrhiza in seedlings but not adult plants of *Plantago lanceolata*. *Plant* and Soil 351, 309–324.
- Svoboda J. & Henry G.H.R. 1987. Succession in marginal Arctic environments. Arctic and Alpine Research 19, 373–384.
- Schweinbacher E., Marcante S. & Erschbamer B. 2010. Alpine species seed longevity in the soil in relation to seed size and shape: a 5-year burial experiment in the central Alps. *Flora* 205, 19–25.
- Väre H., Vestberg M. & Eurola S. 1992. Mycorrhiza and rootassociated fungi in Spitsbergen. *Mycorrhiza 1*, 93–104.
- Whittaker R.J. 1993. Plant population patterns in a glacier foreland succession: pioneer herbs and later-colonizing shrubs. *Ecography 16*, 117–136.
- Yoshitake S., Uchida M., Koizumi H., Kanda H. & Nakatsubo T. 2010. Production of biological soil crusts in the early stage of primary succession on a High Arctic glacier foreland. *New Phytologist 186*, 451–460.